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Removal of crystal violet from aqueous solutions using Arundo donax L.as a cheap adsorbent

Abstract: In the present study, *Arundo donax* L was used as a low-coast biosorbent for the uptake of basic dye crystal violet (CV) from aqueous media. Systematic batch mode studies of adsorption of crystal violet (CV) on *Arundo donax* were carried out as a function of process of parameters includes initial CV concentration, dose of adsorbent, pH, contact time and temperature. The equilibrium adsorption data were analyzed by Langmuir, Freundlich, Temkin and Dubinin–Radushkevich (D-R)isotherm models. The adsorption of CV followed the Langmuir model with a maximum adsorption capacity (q_{max}) of 19.60 mg/g and pseudo second-order kinetics under a specified set of conditions. The thermodynamic parameters like free energy (ΔG°), enthalpy (ΔH°) and entropy (ΔS°) changes for the adsorption of CV ions have been evaluated and it has been found that the reaction was spontaneous and exothermic in nature.

Owing to its rapid adsorption rate and uptake capacity, stem of *Arundo donax* L. seems to be a promising biosorbent for removal of toxic dyes from wastewater.

Keyword: Adsorption, crystal violet, isotherms, kinetics, thermodynamics

1. Introduction:

Dyes are used in large quantities in many industries including textile, leather, cosmetics, paper, printing, plastic, pharmaceuticals, food, etc. to colour their products, which generates wastewater characteristically high in colour¹. Currently, over than 100.000 dyes are commercially available and more than 8 x 10^5 tons are produced annually to supply these industries².

Discharging of dye-containing effluents into hydrosphere is prohibited not only because of their color, which reduces the sunlight penetration into the water, but also because of toxic, carcinogenic and mutagenic nature of their breakdown products³.

As a typical cationic dye, crystal violet (CV) belongs to triphenyl methane group, and is widely applied in coloring paper, temporary hair colorant, dyeing cottons, and wools. CV is harmful by inhalation, ingestion and skin contact, and has also been found to cause cancer and severe eye irritation to human beings^{4,5}.As such, a method for efficiently removing CV molecules from wastewaters is urgently-needed.

Many physical and chemical processes⁶ have been used to remove synthetic dyes from industrial wastewaters such as coagulation-flocculation ⁷, membrane filtration ⁸, fenton process⁹, Photo-catalytic degradation¹⁰, irradiation¹¹, ozonation¹², adsorption¹³.

Among these methods, adsorption has been extensively employed for the removal of dye molecules from different types of water due to its easy operational conditions, low cost, simple design and high efficiency¹⁴. Activated carbon is the most efficient adsorbent used to date, but its high cost limits its applicability Research is currently focusing on the use of low cost commercially available organic materials as viable substitutes for activated carbon; in fact, Lignocellulosic materials exhibit interesting capacities as pollutant adsorbents.

Arundo donax L., belonging to the grass family (Poaceae), is distributed from the Mediterranean region to China and Japan. It is a fast growing perennial grass. Under optimum conditions it can attain growth rate of 0.7 m per week or 10 cm per day, among the fastest growth rate in plants¹⁵. In addition, it contains high amount of lignin and cellulose therefore it is an energy crop, which displays many attractive characteristics for the production of biomass^{16,17}, of pulp and paper¹⁸ and of activated carbons¹⁹. Also, that species could be used as biofiltering material for sewage effluent treatment²⁰ and has been found the capability of accumulate heavy metals to its body²¹.Considering these facts presented above, *Arundo donax* L. could be a non-living biomass with adsorptive properties which can be explored to remove pollutants from water. Our recent findings

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have shown that the stem of *Arundo donax* was able to remove heavy metals²². As an extension to this finding, we explore the possibility of the potential utilization of the stem of *Arundodonax* as means to remove

toxic dyes from aqueous solutions. To the best of our knowledge, there have been no reported studies on the use of *Arundo donax* for the removal of toxic dyes. The objective of this paper is to examine the adsorption of CV on stem of *Arundo donax* This research study is conducted to utilize the *Arundo donax* stem as a potential adsorbent to remove crystal violet from its aqueous solutions. Removal efficiency was tested with a series of batch experiments that varied solution pH, adsorbent dosage, contact time, initial dye concentration and temperature. Removal mechanisms are discussed via FTIR, pH_{pzc} analyses and Boehm titration method. The equilibrium isotherms, kinetics, and thermodynamic parameters were determined to predict the nature of CV adsorption.

2. Materials and Methods:

2.1. Chemicals

All reagents were of analytical reagent grade and used as supplied. Physicochemical characteristics of CV are given in Table 1.

| Dye | Crystal violet | | | |
|-------------------------|--------------------------------|--|--|--|
| C.I. number | 42555 | | | |
| C.I. name | Basic Violet 3 | | | |
| Class | triarylmethane | | | |
| Molecule weight (g/mol) | 407.99 | | | |
| Chemical formula | C25H30C1N3 | | | |
| Water solubility (g/l) | 16 (25°C) | | | |
| λ_{max} (nm) | 590 | | | |
| Synonyms | Basic Violet 3, Gentian Violet | | | |

Molecular structure



Table 1. Physicochemical characteristics of CV



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A stock solution of the dye containing 1000 mg/l was prepared by dissolving the required amount of dye in double distilled water. Solutions of different concentrations used in the further experiments were prepared by diluting the stock solution. The initial pH of the solution was adjusted using 0.1M HCl and 0.1M NaOH solutions.

2.2. Adsorbent preparation:

The *Arundo donax* used in this study was obtained from an agricultural area in El kennar, Jijel, Algeria. The collected biomaterial was extensively washed with tap water to remove soil and dust and then cut to select stems out. The stems were dried by exposure to the sunlight for 3days and subsequently at 80°C for 3 h in a hot air convection oven. The samples were then ground in a mortar and sieved till finally became homogenous powder.

2.3. Adsorbent characterization:

The functional groups of *Arundo donax* were determined using Shimadzu IR 8400s spectrophotometer for Fourier transform infrared spectroscopy (FTIR) analysis using KBr disc method. The point zero charge (pH_{PZC}) was used to determine the point where the density of electrical charge is zero. The solid addition method²³ was used to determine the point zero charge of the adsorbent. The surface acidic and basic function groups were determined by Boehm titration method²⁴.

2.4. Analytical method:

The equilibrium/residual concentrations of CV in solutions were determined with the help of a Shimadzu spectrophotometer(6101-UV–VIS). All measurements were performed at the maximum absorbency visible wave length(485 nm)

2.5. Batch adsorption studies:

Batch adsorption experiments were carried out to understand the effects of solution pH, adsorbent dosage, contact time, initial dye concentrations and temperature.

The dye adsorption was performed in a set of Erlenmeyer flasks where a volume of cationic dye solution of known initial concentration was shaken at the certain agitation speed with a required dose of adsorbent for a specific period of contact time in an orbital shaker. When it reached equilibrium, the solution was centrifuged, the absorbance of supernatants was determined by spectrophotometer, and then, the final concentration was measured.

The adsorption capacity $q_e(mg/g)$ and removal efficiency (%) were calculated as follow:

$$q_e = \frac{(C_0 - C_e)V}{m} \tag{1}$$

% Removal =
$$\frac{(C_0 - C_e)}{C_0} \times 100$$
 (2)

Where C_o and C_e are the initial and final dye concentration(mg/l¹), V is the volume of the dye solution (L) and m is the mass of adsorbent (g).

2.6. Adsorption kinetic and isotherm models:

In order to study the adsorption of CV onto MPO and to interpret the results, experimental data obtained were fitted to different kinetic models such as the pseudo-first-order²⁵, the pseudo-second order²⁶. To establish the rate determining steps for CV

The purpose of the adsorption isotherms is to relate the adsorb at concentration in the

bulk and the adsorbed amount at the interface ²⁸. Thus, isotherm modeling is important for adsorption data interpretation and prediction. The equilibrium characteristics for CV ions adsorption on *Arundo donax* were evaluated by Langmuir, Freundlich, Dubinin-Radushkevich and Temkin isotherms. All kinetic and isotherm models were fitted to experimental data using their non-linear equations (Table 2).

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| | Kinetic models |
|--|---|
| Pseudo-first order | $q_t = q_e \left(1 - e^{-k_t t} \right)$ |
| Pseudo-second order | $q_t = \frac{k_2 q_e^2 t}{1 + k_2 q_2 t}$ |
| Intraparticle diffusion | $q_t = k_{id}t^{0.5} + C$ |
| | Isotherm models |
| Langmuir | $q_e = \frac{q_{\max}K_LC_e}{1 + K_LC_e}$ |
| Freundlich | $q_{\epsilon} = K_F C^{1/n}$ |
| Dabinin-Radushkevich | $\begin{aligned} q_{\epsilon} &= \mathcal{Q}_{m} \exp(-\beta \epsilon^{2}) \\ \varepsilon &= RT \ln(1 + \frac{1}{C_{\epsilon}}) \\ E &= \frac{1}{\sqrt{2\beta_{DR}}} \end{aligned}$ |
| Temkin $q_{\varepsilon} = \frac{RT}{B_{\tau}} \log A_{\tau} C$ | τ. ε |

Table 2. Non-linear equations of kinetic and isotherm models

 $k_1(1/\min) =$ rate constant of pseudo-first order; k_2 (g/mg min) = rate constant of pseudo-second order; $k_{id}(mg/(g \min^{0.5}) =$ intraparticle diffusion rate constant; $q_{max}(mg/g) =$ maximum adsorption capacity; $K_L(L/mg) =$ Langmuir constant; K_f ((mg/g) (L/mg)^{L/n})= Freundlich constant; n= Freundlich exponent; $\beta_{DR} = (g^2/J^2)$ activity coefficient; ϵ (J/g)= Polanyi potential; B_T(J/mol)= Temkin constant; A_T (L/mg)= maximum binding constant; R = universal gas constant (8.314 J/ mol K); T =absolute temperature (K); E(J/mol) = mean free energy of sorption.

3. Results and Discussion:

3.1. Adsorbent characterization

FTIR analysis

FTIR spectrum study was carried out to explain the adsorption mechanism for identifying the presence of functionalities of the *Arundo donax* biomass and analyte.

The FTIR spectrum of *Arundo donax* and CV-loaded *Arundo donax* (Fig. 1 a and b) revealed that upon treatment of adsorbent with CV, the wavelength of the broadband at 3406 cm⁻¹ is decreased to 3396 cm⁻¹indicating that hydroxyl and amino groups areinvolved in bond formation with dye. A significant decrease in wavelength is observed for C=O stretching of the carboxylic acid peak from 1652cm⁻¹ to 1636 cm⁻¹, indicating that carboxylic acid groups play a significant role in the adsorption of CV by *Arundo donax*.

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Fig.1. FTIR spectra of Arundo donax (a) untreated (b) after adsorption with CV

Boehm titration

Boehm titration method was used to determine the surface chemistry and the amount of functional groups such as phenolic, lactonic, and carboxylic groups of *Arundo donax*.

| Table 5. Dochin results of Armitalo uonax | | | | |
|---|-------|--|--|--|
| Active sites | Meq/g | | | |
| Total acidic sites | 1.306 | | | |
| Carboxylic | 0.548 | | | |
| Lactone | 0.130 | | | |
| Phenolic | 0.626 | | | |
| Total basic sites | 0.09 | | | |

| Table 3. | Boehm | results | of Arundo | donax |
|----------|-------|---------|-----------|-------|
|----------|-------|---------|-----------|-------|

Surface functional groups(Table 3) clearly indicates that the total acidic groups are hightly greater (1.306 meq/g), indicating the *Arundo donax* surface is acidic, which resulted from the presence of major acidic groups.

Effect of initial pH

Medium pH plays a significant role in the uptake of adsorbates by biosorbents as the pH would cause the surface charge to alter, thereby affecting the efficiency of biosorbents²⁹.

The pH at the point of zero charge (pH_{PZC}) was used to characterize the properties of the *Arundo donax* surface. Thus, When solution pH <pHpzc, the surface of the adsorbent is predominately positive whereas when solution pH >pHpzc, the adsorbent's surface will be predominately negative. Therefore, the adsorption of cationic species such as CV is favored under the latter condition. The result in Fig. 2 shows that the surface charge of *Arundo donax* at pH 4.9 was zero. Hence, the pH_{pzc} of adsorbent was 4.9.

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Figure 2.Determination of pH_{pzc} of the Arundo donax L.

Fig.3 shows the adsorption of CV as a function of solution pH from 3 to 8. The adsorption capacity of *Arundo donax* was highest (18.42 mg/g) at a pH value of 3; when the pH was increased to 8, the adsorption capacity decreased to 13.28 mg/g but still did not reach equilibrium.



Fig. 3. Effect of pH on the adsorption of CV onto *Arundo donax* ($C_0 = 100 \text{ mg/ l}$, m = 1 g, $T = 25^{\circ}C$ and contact time = 240 min).

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According to the concept of pHpzc, at pH >4.9 where the dye adsorption should be higher due to predominately negatively charged surface thus favoring the adsorption of cationic dye molecules via electrostatic attraction, however experimental data for CV showed otherwise. This disagreement with pH_{pzc} concept merely indicated that the adsorption of CV is not occurring purely by electrostatic interaction, while other attractive forces such as hydrophobic–hydrophobic interaction and hydrogen bonding may be significantly dominant. This behaviour was also observed by^{30, 31}.

Effect of adsorbent dose:

One of the parameters that strongly affect the adsorption removal is the dose of the adsorbents³². The effect of adsorbent dose on the removal of CV by *Arundo donax* at initial concentration ($C_o = 100 \text{ mg/l}$) is illustrated in Fig.4.



Figure 4. Effect of adsorbent dosage on CV removal by *Arundo donax* (C_0 = 100 mg/l, pH = 3, T = 25°C and contact time = 240 min

With increase in adsorbent dose, from 0.5g/l to 4.5 g/l, the amount of adsorbed CV increases from 5.41mg/g to 17.94mg/g and maximum amount removal was observed at adsorbent dose of 4.5 g/l where the quantity removal was 17.94mg/g. This pattern is due to the availability of active sites or surface area on the adsorbent increases with increase of adsorbent dosage facilitates the adsorption of the CV which in turn results in increase of amount removal³³.

Effect of ionic strength:

Generally, most of dying processes utilized in their industries large amounts of salts. Adsorption process can be affected by the presence of some metal ions which are present in wastewaters as ionic strength can affect electrostatic and hydrophobic–hydrophobic interactions. Both electrostatic attraction and repulsion are important for the adsorption. Electrostatic attraction promotes the adsorption; on the contrary, electrostatic repulsion suppresses the adsorption³⁴.

Fig.5 presents the effect of ionic strength on the uptake of CV.

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Fig. 5. Effect of salts concentration for CV adsorption on *Arundo donax* (C₀= 100 mg/l, m=4.5 g/l, contact time = 120 min, pH 3).

It was observed that for both salts, adsorption amount decreased with increase in ionic strength. As NaCl concentration changed from 10 to 50 g/l, the removal of CV decreased by 3.7% (from 17.09% to 13.39%, Fig. 5), while the uptake of CV decreased by 3.89% (from 12.63% to 8.64%, Fig 5) as MgCl₂ concentration varied from 10g/l to 50g/l.

This behavior is usually observed for the removal of dyes and has been reported in many studies³⁵. This reduction in dye uptake at high salt concentration suggested that electrostatic interaction is not the major force of interaction but hydrophobic–hydrophobic interaction and other forces³¹.

Adsorption kinetics:

The adsorption kinetic is one of the most important properties in defining the rate and mechanism of the adsorption process³⁶. Fig. 6 depicts the adsorption of CV dye as a function of contact time using three initial dye concentrations (100mg/l, 200mg/l and 300mg/l).

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Fig. 6. Effect of initial dye concentration and contact time on CV removal By *Arundo donax* (pH 3, m =4.5 g/l and T = 298 K).

The plots show that equilibrium was achieved within 30 min indicating *Arundo donax* was an effective adsorbent for the short contact time with the CV.

The time variation plot indicates that the removal of dye is rapid in initial stages but when it approaches equilibrium, it slows down gradually. The fast adsorption at the initial stage may be due to the fact that a large number of surface sites are available for adsorption but after a lapse of time, the remaining surface sites are difficult to be occupied. This is because of the repulsion between the solute molecules of the solid and bulk phases, thus, make it take long time to reach equilibrium³⁷. When the initial dye concentration increased from 100mg/l to 300mg/l the adsorption capacity increased from 17.28 mg/g to 58.21 mg/g. This is due to increase in the driving force of the concentration gradient, as an increase in the initial concentration. The concentration provides an important driving force to overcome all mass transfer resistance of the dye between the aqueous and solid phases. Hence a higher initial concentration of dye will enhance the adsorption process³⁸. The kinetic parameters obtained for CV adsorption are listed in Table 4.

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| C ₀ (mg/l) | qe,exp(mg/l) | Pseudo-first order | | Pseudo-second order | | Intraparticle diffusion | | | | |
|-----------------------|--------------|--------------------|-------|---------------------|----------------|-------------------------|----------------|-----------------|-------|----------------|
| | | \mathbf{k}_1 | q_e | R ² | k ₂ | q_e | R ² | K _{id} | qe | R ² |
| 100 | 16.80 | 0.213 | 3.20 | 0.804 | 0.037 | 16.66 | 0.999 | 1.21 | 6.05 | 0.78 |
| 200 | 38.51 | 0.089 | 90.9 | 0.834 | 0.033 | 38.46 | 1 | 5.19 | 34.81 | 0.64 |
| 300 | 58.21 | 0.071 | 16.44 | 0.821 | 0.015 | 58.82 | 0.999 | 7.75 | 73.52 | 0.847 |
| | | | | | | | | | | |

Table 4 Parameters of the kinetic models for CV adsorption by Arundo donax.

The results show (Table 4, Fig.7a,b and c) for pseudo-first kinetic model that the values of R² were low and the experimental $q_{e,exp}$ values do not agree well with the calculated values. This shows that the adsorption of CV onto *Arundo donax* is not first-order kinetic, indicating that the adsorption was not diffusion-controlled and adsorption was not preceded by diffusion through a boundary³⁹. This is confirmed by results obtained with the intra-particle model. In this model, the plot of q_t versus $t^{1/2}$ were not linear over the whole time range. Furthermore, it may be seen that the intra-particle diffusion of CV dye occurred in 2 stages. The first straight portion indicates that boundary layer diffusion probably limited CV adsorption⁴⁰ and the second linear portion is attributed to intra-particle diffusion. The correlation coefficient values for the pseudo-second order rate equation (R² = 0.9993–0.9999)(Table 4, Fig. 7b) were found to be higher than the pseudo-first order rate equation (R² = 0.9653-0.9451) and the $q_{e,calc}$ and $q_{e,exp}$ values were in close agreement with each other for pseudo-second order model, which conformed that the adsorption process followed the pseudo-second order kinetic model⁴¹.



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Fig. 7. Kinetics plots for the adsorption of CV onto *Arundo donax*: (a) pseudo-first order, (b): pseudo-second order, (c): intraparticle diffusion

Adsorption isotherms:

The purpose of the adsorption isotherms is to relate the adsorbate concentration in the bulk and the adsorbed amount at the the interface⁴². The non-linear isotherm models of Langmuir, Freundlich, Dubinin-Radushkevich(D-R) and Temkin were adjusted to experimental data and they are shown in Fig. 8.

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Fig. 8. Isotherms of CV) adsorption by Arundodonax as a function of temperatures: (A): 25° C, (B): 35° C and (C): 45° C

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The Langmuir model⁴³ is based on the assumption that the maximum adsorption occurs when a saturated monolayer of solute molecules is present on the adsorbent surface, the energy of adsorption is constant and there is no migration of adsorbate molecules in the surface plane.

The Freundlich isotherm model⁴⁴ is the earliest known relationship describing the non-ideal and reversible adsorption, which can be applied to multilayer adsorption, on the basis of an assumption concerning the energetic surface heterogeneity. The Freundlich adsorption capacity magnitude (n) is useful to assess the favorability of the adsorption. The n values between 2-10 indicate high adsorption capacity, whereas n values between 1-2 indicate moderate adsorption capacity, and n values lower than 1 suggest a poor adsorption capacity.

Temkin⁴⁵ considered the effects of some indirect adsorbate/adsorbate interactions on adsorption isotherms and suggested that because of these interactions the heat of adsorption of all the molecules in the layer would decrease linearly with coverage. The Dubinin-Redushkevich (D-R) model⁴⁶ is an empirical equation which is applied to distinguish between physical and chemical adsorption. The β_{D-R} is the activity coefficient related to mean adsorption energy and E is the mean free energy, which can be calculated by using β_{D-R} values. For magnitude of E between 8 KJ/mol and 16KJ/mol, the adsorption process followed chemical ion exchange, and values of E below 8 KJmol⁻¹ were the characteristic of physical adsorption process⁴⁷. The parameters obtained for each model and their correlation coefficients are displayed in Table 5. High R² are derived by fitting experimental data into the Langmuir isotherm model (R² > 0.99) as compared with the Freundlich (R² >0.63-0.87), Dubinin-Radushkevich (R² >0.85-0.97) and Temkin (R² > 0.64-0.86) isotherm models.

| Isotherm models | 25°C | 35°C | 45°C | |
|----------------------------|--------|--------|-------|--|
| Langmuir | | | | |
| $q_{max}(mg/g)$ | 19.60 | 15.87 | 12.65 | |
| K _L (l/mg) | 0.228 | 0.205 | 0.439 | |
| R ² | 0.999 | 0.999 | 0.998 | |
| Frendlich | | | | |
| $K_{ m F}$ | 3.13 | 3.10 | 3.56 | |
| Ν | 3.01 | 3.23 | 4.18 | |
| R ² | 0.874 | 0.800 | 0.630 | |
| Dubinin-Raduschkevich(D-R) | | | | |
| $q_{max}(mg/g)$ | 15.81 | 13.31 | 13.14 | |
| B _{DR} | 7.10-7 | 8.10-7 | 10-6 | |
| E(KJ/mol) | 0.85 | 0.79 | 0.71 | |
| R ² | 0.808 | 0.857 | 0.973 | |
| Temkin | | | | |
| A _T (l/mg) | 5.023 | 4.172 | 30.03 | |
| B _T | 2.778 | 2.299 | 1.454 | |
| R ² | 0.811 | 0.862 | 0.640 | |

Table 5 Parameters for adsorption of CV on Arundo donax for various isotherm models

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The Arundo donax showed maximum monolayer adsorption capacity (q_{max}) for CV of 19.60 mg/g. This finding was compared to other studies on the adsorption of CV on different adsorbents (Table 6) The most important parameter to

compared to other studies on the adsorption of CV on different adsorbents (Table 6). The most important parameter to compare is the Langmuir q_{max} value since it is a measure of adsorption capacity of the adsorbent. The value of q_{max} in this study is larger than those in most of previous works. This suggests that CV could be easily adsorbed on *Arundo donax*.

| Adsorbents | $q_{max}(mg/g)$ | Keierence | |
|------------------------|-----------------|------------------------------|---|
| Arundo donax | 19.60 | This work | |
| Orange peel | 11.50 | Annadurai et al. (2002) | |
| Banana peel | 12.20 | Annadurai et al. (2002) | |
| peat | 8.20 | Zehra et al. (2016) | |
| Calotropisprocera leaf | 4.10 | Ali and Muhammad (2008) | |
| Citrulluslanatus rind | 12.10 | Suyamboo and Perumal (2012), | |
| Nanomagnetic iron oxid | 12.70 | Hamidzadeh et al. (2015) | |
| Chitosan | 19.23 | Shouman et al. (2012) | |
| | | | 1 |

Table 6 Maximum adsorption capacities of GV from aqueous media using various adsorbents.

Thermodynamic studies:

The effect of a change in temperature on the CV-Arundo donax adsorption system was studied to determine the thermodynamic parameters and to investigate the nature of the process. The change in Gibbs free energies(ΔG°) was calculated with equation3.

 ΔH° and ΔS° were calculated from the slope and intercept of the plot of $\ln K_{d}$ versus 1/T using Van't Hoff equation (data not shown):

$$AD = -RTLnK_{D} \tag{3}$$

$$LnK_{d} = \frac{\Delta S^{0}}{R} - \frac{\Delta H^{0}}{RT}$$
(4)

Where R is the universal gas constant (8.314 J mol⁻¹K⁻¹), T is the temperature of the medium (K) and K_d is the distribution coefficient. It is determined from:

$$K_d = \frac{q_e}{C_e} \tag{5}$$

All the thermodynamic parameters are listed in Table 7. The negative value of ΔH° shows that the adsorption of CV on *Arundo donax* is exothermic process⁴⁸. The negative value of ΔS° indicates the decreased disorder and randomness at the solid solution interface of CV with adsorbent. The negative values of ΔG° indicate the adsorption is favorable and spontaneous. ΔG° values decreases with increase in temperature. The decrease in the value of ΔG° with temperature further showed the decrease in feasibility of adsorption at elevated temperatures. The change in free energy for physisorption is between -20 and 0 KJ/mol, but chemisorption is in a range of -80 to -400 KJ/mol⁴⁹. The values of ΔG° obtained in this study are within the ranges of -20and 0 KJ/mol, indicating that the physisorption is the dominating mechanism.



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| T(K) | ΔG^0 (kJ mol ⁻¹) | ΔS^{0} (J mol ⁻¹ K ⁻¹) | ΔH^0 (kJ mol ⁻¹) | R ² |
|------|--------------------------------------|---|--------------------------------------|-----------------------|
| 298 | -07.31 | | | |
| 308 | -08.10 | -89.08 | -18.98 | 0.999 |
| 318 | -09.41 | _ | | |

Table 7 Thermodynamic parameters of the CV adsorption onto Arundo donax at different temperatures

4. CONCLUSION:

The results obtained in this study indicate that the *Arundo donax* can be successfully used for the removal of hazardous dye, crystal violet from aqueous solutions. The advantages of this biosorbent are low cost, high availability, and good adsorption capacity. The batch adsorption process is found depend upon pH, temperature, adsorbent dosage and initial dye concentration. The adsorption process was not severely affected by high ionic strength.

The Langmuir isotherm model best represented the adsorption process which predicted q_{max} at 19.60 mg/g at 25°C. Kinetics of adsoption of CV was fast and followed the pseudo-second-order. Thermodynamic studies showed that adsorption CV-*Arundo donax* system is spontaneous and exothermic. Thus, it can be concluded that the biomaterial *Arundo donax* biomass can be used as excellent sorbent for the removal of dyes from wastewater.

REFERENCES:

- 1. S.S.Azhar, A.G. Liew, D.Suhardy, K.F. Hafiz, M.D. Iatim, Am, J, Appl, Sci, 200(2005) 1499-1503
- 2. M.X. Wang, Q.L. Zhang, S.J. Yao, Chem, Eng, J, 259(2015) 837-844
- 3. S.Wang, Y. Boyjoo, A.Choueib, Z.H.U. Zhu, Removal of dyes from aqueous solution using fly ash and red mud, Water Res., 39(2005) 129-138
- 4. Y.Lin, X. He, G. Han, Q.Tian, W. Hu, J, Environ, Sci,23(2011) 2055-2062
- 5. A.Saeed, M. Sharif, M.Iqbal, Bioresour, Technol, 179(2010) 564-572
- 6. U.A.Guler, M.Ersan, E.Tuncel, F.Dügenci, Process, Saf, Environ, Prot, 99(2016) 194-206
- 7. A.Szyguła, E.Guibal, M.ArinoPalacín, M. Ruiz, A.M.Sastre, J, Environ, Manag, 90(2009) 2979-2986
- 8. E.Alventosa-deLara, S.Barredo-Damas, M.I.Alcaina-Miranda, M.I.Iborra-Clar, J, Hazard, Mater, 209-210(2012) 492-504
- 9. M.Karatas, Y.A. Argun, M.E.Argun, J,Ind,Eng, Chem, 18(2012) 1058-1062
- 10. T.A.Saleh, V.K. Gupta, J, Colloid, Interfac, eSci, 371(2012) 101-106
- 11. J.Paul, K.P.Rawat, K.S.S. Sarma, S.Sabharwal, Appl, Radiat, Isot, 69(2011) 982-987
- 12. G.Moussavi, M.Mahmoudi, Chem, Eng, J, 152(2009) 1-7
- 13. F.Krika, O.F.Benlahbib, Desalin, Water, Treat, 53(2015) 3711-3721
- 14. X.S.Wang, J.P. Chen, Clean, Soil, Air, Water, 37(2009) 793-798
- 15. G.Bell, In Brock, J. H., Wade, M., Pysek, P., and Green, D. (Eds.): Plant Invasions: Studies from North America and Europe. Blackhuys Publishers, Leiden, The Netherlands, pp. 103-113(1997)

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INTERNATIONAL JOURNAL OF SCIENCE AND TECHNOLOGY

- 16. L.G.Angelini, L.Ceccarini, E.Bonari, Eur, J,Agron, 22(2005) 375-389
- 17. I.Lewandowski, J.M.O.Scurlock, E.Lindvall, M.Christou, Biomass, Bioenergy, 25(2003) 335-361
- 18. A.A.Shatalov, H. Pereira, Bioresour, Technol, 96(2005) 865-872
- 19. T. Vernersson, P.R.Bonelli, E.G.Cerrella, A.L.Cukierman, Bioresour, Technol, 8(2002) 95-104
- 20. G.Mavrogianopoulos, V.Vogli, S. Kyritsis, Bioresour, Technol, 82(2002) 103-107
- 21. H.Deng, Z.H.Ye, M.H. Wong, Environ, Pollut, 132(2004) 29-40
- 22. H.L.Song, L.Liang, K.Y. Yang, Chem, Eng., Res, Des, (2014), doi:10.1016/j.cherd.2014.04.027
- 23. L.S.Balistrieri, J.W. Murray, Am, J,Sci,281(1981) 788-806
- 24. S.L.Goertzen, K.D.Theriault, A.M. Oickle, A.C.Tarasuk, H.A Andreas, Carbon, 48(2010) 1252-1261
- 25. S.Lagergren, Handl, 24(1898) 1-39
- 26. Y.S.Ho, M.C. Kay, Process, Biochem, 34(1999) 451-465
- 27. W.J.Weber, J.C.Morris, J,Sanit,Eng,Div,89(1963) 31-59
- 28. J.Eastoe, J.S. Dalton, Adv, J, Colloid, Interface, Sci,85(2000) 103-144
- 29. L.B.L.Lim, N.Priyantha, C.H.Ing, M.K.Dahri, D.TB.Tennakoon, T.Zehra, M.Suklueng, Desalin, Water, Treat, (2013), doi: 10.1080/19443994.2013.852136
- 30. Y.S.Al-Degs, M.I. El-Barghouthi, A.H. El-Sheikh, G.M. Walker, Dyes and Pigments, 77(2008) 16-23
- 31. M.K.Dahri, M.R.R.Kooh, L.B.L Lim, Alexandria, Eng, J, (2015), doi:10.1016/j.aej.2015.07.005
- 32. M.M.Abd El- Latif, A.M. Ibrahim, Desalin, Water, Treat, 6(2009) 252-268
- 33. S.Kazemi, P.Biparva, E.Ashtiani, Ecol, Eng, 88(2016) 82-89
- 34. Y.Hu, T.Guo, X. Ye, Q. Li, M.Guo, H. Liu, Z. Wu, Chem, Eng, J, 228(2013) 392-397
- 35. M.K.Dahri, M.R.R.Kooh, L.B.L. Lim, J, Environ, Chem, Eng, 2(2014) 1434-1444
- 36. M.Rathod, K.Mody, S.Basha, J, Clean, Prod, (2014), doi: 10.1016/j.jclepro.2014.03.064
- 37. Ö.Dülger, F.Turak, K.Turhan, M.Özgür, Anal, Chem, (2013), Article ID 210470, 9 pages
- 38. S.Sener, Chem, Eng, J,138(2008) 207-214
- 39. N.Dizge, C.Aydiner, E.Demirbas, M.Kobya, S. Kara, J, Hazard, Mater, 150(2008) 737-746
- 40. H.Mittal, S.B. Mishra, Polym, 101(2014) 1255-1264
- 41. J. Eastoe, J. Dalton, Adv, J, Colloid, Inter, Sci,85(2000) 103-144
- 42. M.N.Idris, M.A.Ahmad, Int, J, Basic, App, Sci,11(2011) 38-43
- 43. I.Langmuir, J, Am, Chem, Soc, 40(1918) 1361-1368
- 44. H.Freundlich, Phys, Chem, Soc, 40(1906) 1361-1368
- 45. M.J.Temkin, V.Pyzhev, Acta, Physiochem, 2(1940) 217-222
- 46. M.M.Dubinin, L.V.Radushkevich, Chem, Zent,1(1947) 875-889
- 47. A.P.Meneghel, A.C.J.Gonçalves, F.RubioD.C.Dragunski, C.A.Lindino, LStrey, Water, Air, Soil,Pollut,224(2013) 1383-1396
- 48. S. Arivoli, B.R.Venkatraman, T.Rajachandrasekar, Res, J,Chem, Environ, 11(2007) 70-78.
- 49. M.J.Jaycock, G.D.Parfitt, Chemistry of Interfaces, Ellis Horwood Ltd., Onichester (1981).

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